

# Maranthaceae overabundance decreases richness and abundance of regenerating woody plants in natural gaps

## A superabundância de Maranthaceae reduz a riqueza e abundância de espécies arbustivo-arbóreas em regeneração em clareiras naturais

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### Abstract

Natural gaps are known to be one of the most common disturbance types in tropical forests. Maranthaceae is one of the most important herb families and some studies on gaps address the influence of herbs on the regeneration of shrubs and trees. Maranthaceae are common and sometimes dominant herbs in tropical forest gaps. In this study we investigated the impact of Maranthaceae overabundance on the regeneration of woody species in natural gaps in a bamboo-dominated (*Guadua weberbaueri* Pilg.; Poaceae) forest fragment in southwestern Amazonia. We sampled 20 natural gaps, where Maranthaceae and woody plants with height  $\geq 1$  m and DBH  $\leq 10$  cm were counted, measured and identified. We used NMDS for species composition analyses and a GLM to test the influence of Maranthaceae overabundance on woody species richness and abundance. We also measured the similarity of woody species richness and abundance among Maranthaceae dominated and non-dominated gaps. The NMDS shows a distinction between Maranthaceae dominated and non-dominated gaps. The overabundance of Maranthaceae reduced significantly the richness and abundance of woody species. Species composition was also dissimilar among gaps and was influenced by Maranthaceae overabundance.

**Keywords:** natural disturbances, herbaceous, plant succession.

### Resumo

Clareiras naturais são um dos tipos de distúrbios mais estudados em florestas tropicais. Maranthaceae é uma das famílias mais importantes de ervas terrestres e alguns estudos têm mostrado a dominância de espécies dessa família em relação a outras espécies herbáceas em clareiras naturais. Nesse contexto, o objetivo deste trabalho foi verificar se existe influência da superabundância de Maranthaceae sobre a regeneração de espécies arbustivo-arbóreas em clareiras naturais em um remanescente de floresta aberta dominada por bambu (*Guadua weberbaueri* Pilg.; Poaceae) no sudoeste da Amazônia. Foi amostrada a regeneração  $\geq 1$  cm de diâmetro e os indivíduos de Maranthaceae em 20 clareiras formadas pela queda de árvores do dossel. Foi realizada uma análise não métrica de escalas multidimensionais (NMDS) para análise de composição de espécies e foram aplicados modelos lineares generalizados (GLM) para testar o efeito da superabundância de Maranthaceae sobre a riqueza e abundância de árvores e arbustos e da similaridade entre as clareiras dominadas e não dominadas por essas herbáceas. A NMDS realizada mostrou separação das clareiras que foram dominadas por indivíduos de Maranthaceae daquelas não dominadas. A superabundância de Maranthaceae reduziu significativamente a abundância e riqueza de espécies e tornou os ambientes dominados dissimilares em termos de composição de espécies.

**Palavras-chave:** distúrbios naturais, herbáceas, sucessão vegetal.

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## Introduction

The mechanisms that maintain high tree diversity in tropical forests have long drawn the interest of ecologists. The study of these mechanisms led to the formulation of several hypotheses aimed at explaining the coexistence of a large number of species in a small area (Connell, 1978; Denslow, 1980; Hubbell *et al.*, 1999; Wright, 2002). Natural gaps, formed by the fall of one or more canopy trees, contribute to environmental heterogeneity because they modify the microclimate and alter plant dominance in these sites (Veblen, 1989; Denslow, 1987). The formation of natural gaps is the most common and best-studied disturbance in forests (Schliemann and Bockheim, 2011). The main differences between gaps and areas with closed canopy are the increase in light incidence and changes in light quality.

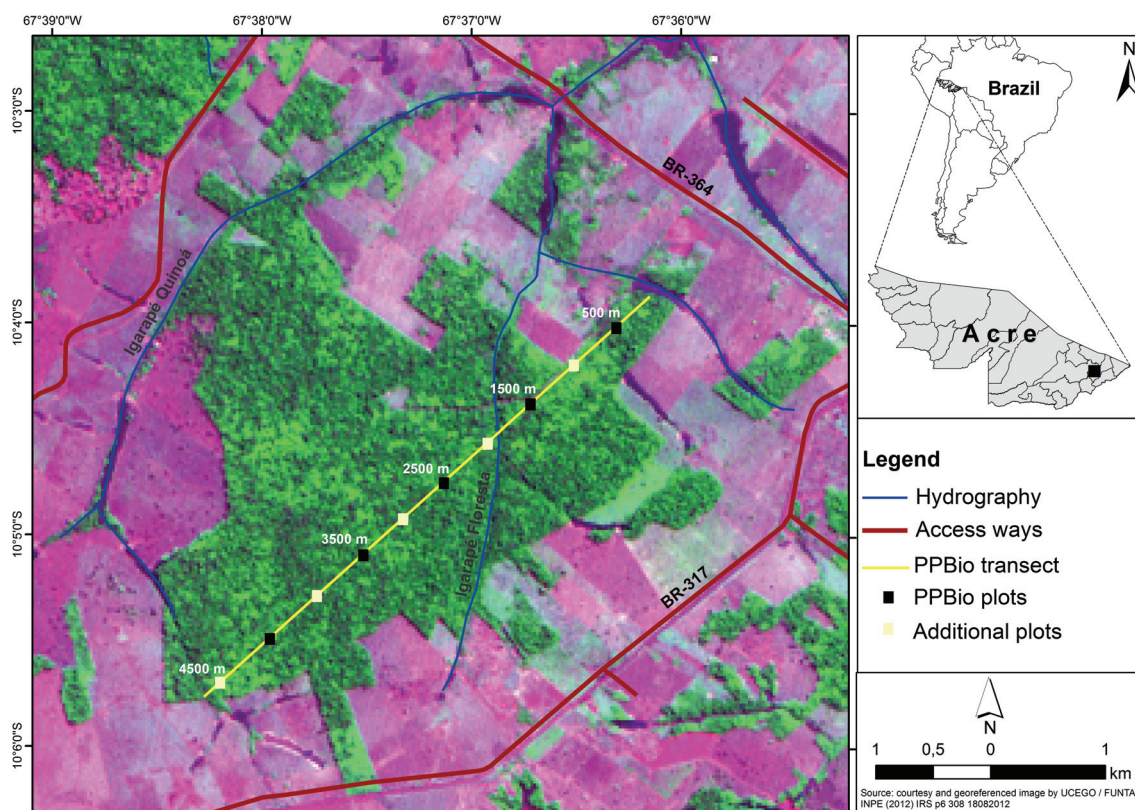
In addition, resource availability, such as water and nutrients, can be temporarily increased in gaps (Brandani *et al.*, 1988; Denslow, 1980, 1987). Invasion and dominance by exotic species are well studied in ecology. However, little attention is given to the local expanding native species that negatively affect other native species, which is known as overabundance (Garrot *et al.*, 1993), in particular terrestrial herbs, on which there are only few studies in the tropics (Costa, 2004). Good examples of overabundant native species are bamboos (*Guadua* spp.; Lima *et al.*, 2012) and ferns (*Pteridium* spp.; Schneider, 2006; Ribeiro *et al.*, 2013), which influence forest structure. Gentry and Emmons (1987) have identified Maranthaceae as one of the most important families among the understory terrestrial herbs of Neotropical forests. Poulsen and Henrik

(1991) and Costa (2004) in their studies have shown the dominance of the species of this family in Amazonian forests of Ecuador and Brazil. Hence, the objective of the present study was to investigate the regeneration of shrubby-arboreal species in natural gaps in a forest remnant dominated by bamboos (*Guadua weberbaueri* Pilg.) in southwestern Amazon. We aimed at testing whether the abundance of Maranthaceae affects the regeneration of native species in natural gaps.

## Material and methods

### Study area

We carried out the present study at the Catuaba Experimental Farm (CEF), a forest remnant of approximately 1,200 ha, located in Senador Guiomard, state of Acre (Figure 1), northern Brazil



**Figure 1.** Study area. PPBio plots following the RAPELD protocol, established at the Catuaba Experimental Farm, southwestern Amazonia, Brazil. Each black square indicates the initial position of a 250-m long plot and the yellow squares indicate additional trails marked.

(10°04' S and 67°37' W). The local topography is slightly corrugated, with a predominance of latosols and argisols (Acre, 2006); the horizons A and B are predominantly composed of sand (62 and 47%, respectively); pH approximately 4.0 (Sousa *et al.*, 2008). The local altitude is 214 m a.s.l. and the studied forest remnant is located 0.8–7.4 km away from neighbor remnants. It is covered by open rainforest dominated by bamboos (*G. weberbaueri* Pilg.). The local dominant tree species are *Hevea brasiliensis* (WILLD. EX. A. JUSS.) MÜLL. ARG. (Euphobiaceae), *Bertholletia excelsa* HUMB. & BONPL. (Lecythidaceae), *Tetragastris altissima* (AUBL.) SWART (Burseraceae), and *Carapa guianensis* AUBL. (Meliaceae). The canopy height varies from 20 m to 40 m, with emergent trees reaching up to 45 m (Silveira, 2005). The average annual rainfall is 1,958 mm and the average annual temperature is 25 °C (Duarte, 2006).

For mapping the natural gaps, we walked 10 km of trails, following Runkle (1992), due to standardization for location of natural gaps. We considered all gaps formed by the fall of one or more trees, as well as of part of a tree, in a plot of the Research Program in Biodiversity (PPBio – Programa de Pesquisas em Biodiversidade) established at CEF, following the RAPELD (rapid assessments and long-term ecological research) protocol (Magnusson *et al.*, 2005). According to this protocol, a 5 km transect was created, with 10 perpendicular plots (250 m by 20 m width on each side, following the same topographic height), summing 10 plots along the 5 km transect, 500 m apart from each other. In the present study we considered only gaps with a minimum area of 100 m<sup>2</sup>.

### Data sampling

We used the operational definition of gap by Runkle (1981), which includes the area of soil under the canopy opening, extending to the base of adjacent

trees. In total, we sampled 20 gaps. We measured the area of each gap with a GPS Garmin 78 (precision of ±4 m), and the percentage of canopy opening with a hemispherical photograph.

In each gap, we installed eight subplots (2 m x 4 m) to represent its heterogeneity (Brandani *et al.*, 1988). Inside these subplots, we sampled all shrubby-arboreal plants with height ≥ 1 m and DBH ≤ 10 cm. We marked each individual plant with a numbered plate. We identified the species in the field with the help of a parataxonomist. In addition, we collected samples from each species for identification by comparison with vouchers deposited in the herbarium of the Federal University of Acre (UFAC – Universidade Federal do Acre) and by consulting the specialized literature. Later, the samples were deposited in the collection of the Laboratory of Botany and Plant Ecology of UFAC, Rio Branco, Acre, northern Brazil.

As an operational definition, we considered a Maranthaceae dominated gap when there were one or more Maranthaceae individuals per square meter, since they achieve up to 2m height and long leaves, physically occupying most of the area.

To test for an influence of Maranthaceae abundance on tree regeneration, we counted all individuals in each subplot. Due to the form of growth of these species, we considered each clump as one individual.

### Data analysis

To test for an influence of Maranthaceae abundance on the composition of woody species, we used a non-metric multidimensional scaling analysis (NMDS), using species abundance data and 10,000 permutations. To build the dissimilarity matrix, we used the Bray-Curtis distance.

We used a generalized linear model (GLM) to estimate the influence of Maranthaceae abundance on woody species, by plotting it with the first axis of the NMDS (Maranthaceae as

independent variable). Maranthaceae overabundance effect on woody species richness and abundance was also tested with GLM. All analyses were carried out in R environment (R core Team, 2013), using Vegan package (Oksanen *et al.*, 2013).

## Results

The area of the studied gaps varied from 108 to 1,413 m<sup>2</sup>, with an average value of 521 ± 347 m<sup>2</sup> (average ± Standard Deviation) and a median value of 353 m<sup>2</sup>. The average percentage of canopy opening was 49%; small gaps showed an average canopy opening of 50%, intermediate gaps of 47%, and large gaps of 52%.

We recorded 1,218 shrubby-arboreal individual plants of 145 species, 111 genera, and 45 families (Table 1). The most abundant species were *Piper* sp. (Piperaceae; 16.4%), *Faramea capillipes* MÜLL. ARG. (Rubiaceae; 5.7%), *Tachigali setifera* (DUCKE) ZARUCCHI & HEREND. (Fabaceae; 3.4%), *Brosimum guianense* (AUBL.) HUBER (Moraceae; 3.0%), *Inga* sp. (Fabaceae; 2.5%), *Siparuna guianensis* AUBL. (Siparunaceae; 2.5%), *Guarea* sp. (Meliaceae; 2.3%), *Amphirrhox* sp. (Violaceae; 2.1%), *Eugenia* sp. (Myrtaceae; 2.1%), *Amphiodon effusus* HUBER (Fabaceae; 2.0%), and *Pseudolmedia laevis* (RUIZ & PAV.) J.F.MACBR. (Moraceae; 2.0%). The NMDS (stress = 0.16) showed a separation between the gaps dominated by Maranthaceae from those not dominated by this family. This result suggests different species composition between gaps with and without dominance of Maranthaceae (Figure 2).

Maranthaceae abundance was marginally correlated with the dissimilarity among gaps, represented by NMDS first axis ( $p = 0.07$ , Figure 3). Gaps that are not dominated by Maranthaceae tended to be more dissimilar among each another.

Where there was high dominance of Maranthaceae, there was a significant decrease in the abundance of woody plants, which suggests a negative

**Table 1.** List of species found in natural gaps, in Catuaba Experimental Farm, southwestern Amazonia, Brazil, with their abundance and percentage of gaps in which they were recorded.

Family	Species	Habit	Gaps occupied (%)	Abundance	Voucher A.S. Maranhão*
Acanthaceae	<i>Justicia</i> sp.	Shrub	5	4	6209
Achariaceae	<i>Lindackeria paludosa</i> (BENTH.) GILG	Shrub	5	1	6249
Annonaceae	<i>Anaxagorea brevipes</i> BENTH.	Tree	5	1	6243
	<i>Duguetia hadrantha</i> (DIELS) R.E.FR.	Tree	5	1	6231
	<i>Guatteria olivacea</i> R.E.FR.	Tree	20	6	6223
	<i>Malmea</i> sp.	Shrub	25	10	6232
	<i>Xylopia</i> sp.	Tree	5	1	
Apocynaceae	<i>Aspidosperma parvifolium</i> A.DC.	Tree	60	16	6199
	<i>Aspidosperma rigidum</i> RUSBY	Tree	5	4	6200
	<i>Geissospermum sericeum</i> MIERS	Tree	5	1	6307
	<i>Himatanthus sucuuba</i> (SPRUCE EX MÜLL.ARG.) WOODSON	Tree	10	6	
Bignoniaceae	<i>Handroanthus serratifolius</i> (A.H.GENTRY) S.GROSE	Tree	5	1	
Boraginaceae	<i>Cordia alliodora</i> (RUIZ & PAV.) CHAM.	Tree	10	2	6295
	<i>Cordia nodosa</i> LAM.	Tree	25	6	6296
Burseraceae	<i>Protium subserratum</i> (ENGL.) ENGL.	Tree	10	2	6313
	<i>Tetragastris altissima</i> (AUBL.) SWART	Tree	35	16	6293
Cannabaceae	<i>Celtis schippii</i> TREL. ex STANDL.	Tree	45	20	
Caricaceae	<i>Jacaratia spinosa</i> (AUBL.) A.DC.	Tree	10	3	6260
Celastraceae	<i>Cheiloclinium</i> sp. MIERS	Tree	5	1	
	<i>Salacia</i> sp.		5	1	6312
Chrysobalanaceae	<i>Hirtella racemosa</i> LAM var. <i>racemosa</i> LAM.	Tree	55	17	
	<i>Licania caudata</i> PRANCE	Tree	10	2	6304
Clusiaceae	<i>Vismia guianensis</i> (AUBL.) CHOISY	Tree	5	1	6208
Combretaceae	<i>Terminalia amazonia</i> (J.F.GMEL.) EXELL	Tree	5	3	6252
Ebenaceae	<i>Diospyros</i> sp.	Tree	20	4	6280
Elaeocarpaceae	<i>Sloanea guianensis</i> (AUBL.) BENTH.	Tree	40	9	6242
Euphorbiaceae	<i>Acalypha macrostachya</i> JACQ.	Shrub	5	1	6214
	<i>Conceveiba guianensis</i> AUBL.	Tree	5	1	6206
	<i>Hevea brasiliensis</i> (WILLD. EX A.JUSS.) MÜLL.ARG.	Tree	20	8	
	<i>Nealchornea</i> sp.	Tree	30	14	6319
Fabaceae:					
Caesalpinioideae	<i>Apuleia leiocarpa</i> (VOGEL) J.F.MACBR.	Tree	10	2	6215
	<i>Barnebydendron riedelii</i> (TUL.) J.H.KIRKBR.	Tree	5	1	6248
	<i>Dialium guianense</i> (AUBL.) SANDWICH	Tree	40	9	6269
	<i>Hymenaea parvifolia</i> HUBER	Tree	5	1	6203
	<i>Poeppigia procera</i> C.PRESL	Tree	5	1	6236
	<i>Tachigali setifera</i> (DUCKE) ZARUCCHI & HEREND.	Tree	50	41	6233
Fabaceae:					
Faboideae	<i>Amphiodon effusus</i> HUBER	Tree	50	24	6281

Table 1. Continuation.

Family	Species	Habit	Gaps occupied (%)	Abundance	Voucher A.S. Maranhão*
	<i>Bowdichia</i> sp.	Tree	10	2	6198
	<i>Erythrina</i> sp.	Tree	15	4	
	<i>Myroxylon balsamum</i> (L.) HARMS	Tree	5	1	6286
	<i>Ormosia</i> sp.	Tree	15	3	6278
	<i>Pterocarpus amazonum</i> (BENTH.) AMSHOFF	Tree	15	4	6210
	<i>Swartzia oraria</i> R.S.COWAN	Tree	25	5	6265
	<i>Swartzia</i> sp.	Tree	5	2	6266
	<i>Vatairea fusca</i> (DUCKE) DUCKE	Tree	5	1	6211
Fabaceae: Mimosoideae	<i>Abarema laeta</i> (BENTH.) BARNEBY & J.W.GRIMES	Tree	5	1	6283
	<i>Anadenanthera</i> sp.		5	1	
	<i>Enterolobium maximum</i> DUCKE	Tree	5	1	6193
	<i>Inga cylindrica</i> (VELL.) MART.	Tree	5	1	6235
	<i>Inga</i> sp. 1	Tree	5	1	
	<i>Inga</i> sp. 2	Tree	40	11	6320
	<i>Inga</i> sp. 3	Tree	15	4	6196
	<i>Inga</i> sp. 4	Tree	55	31	
	<i>Stryphnodendron guianense</i> (AUBL.) BENTH.	Tree	5	1	6258
	Mimosoideae 1		5	1	
	Mimosoideae 2		5	1	
	Mimosoideae 3	Tree	30	9	
Lauraceae	<i>Endlicheria</i> sp.	Tree	5	1	6221
	<i>Mezilaureus sprucei</i> (MEISN.) TAUB. EX MEZ	Tree	20	4	6212
	<i>Ocotea oblonga</i> (MEISN.) MEZ	Tree	5	1	6297
	<i>Ocotea</i> sp. 2	Tree	30	7	6254
	<i>Ocotea</i> sp.	Tree	15	3	6317
Lecythidaceae	<i>Couratari guianensis</i> AUBL.	Tree	40	15	6234
	<i>Eschweilera coriacea</i> (DC.) S.A.MORI	Tree	15	3	
	<i>Eschweilera truncata</i> A.C.SM.	Tree	20	4	6225
	<i>Gustavia augusta</i> L.	Tree	40	17	6253
Malvaceae	<i>Herrania nitida</i> (POEPP.) R.E.SCHULT.	Tree	10	2	6279
	<i>Pseudobombax munguba</i> (MART. & ZUCC.) DUGAND	Tree	10	2	6207
	<i>Quararibea guianensis</i> AUBL.	Tree	70	21	6267
	<i>Sterculia</i> sp.	Tree	5	1	6244
	<i>Theobroma obovatum</i> KLOTZSCH EX BERNOULLI	Tree	45	19	6230
	<i>Huberodendron swietenoides</i> (GLEASON) DUCKE	Tree	10	2	6213
Melastomataceae	<i>Miconia affinis</i> DC.	Shrub	10	3	6259
	<i>Miconia</i> sp.		5	4	
	<i>Mouriri myrtifolia</i> SPRUCE EX TRIANA	Tree	20	7	6216

Table 1. Continuation.

Family	Species	Habit	Gaps occupied (%)	Abundance	Voucher A.S. Maranhão*
Meliaceae	<i>Guarea</i> sp.	Tree	65	28	6322
	<i>Trichilia pleeana</i> (A. JUSS.) C. DC.	Tree	15	3	6247
Moraceae	<i>Brosimum guianense</i> (AUBL.) HUBER	Tree	70	37	6264
	<i>Brosimum lactescens</i> (S. MOORE) C. C. BERG	Tree	5	1	6224
	<i>Castilla ulei</i> WARB.	Tree	5	1	
	<i>Clarisia ilicifolia</i> (SPRENG.) LANJ. & ROSSBERG	Tree	45	19	6276
	<i>Clarisia racemosa</i> RUIZ & PAV.	Tree	15	4	6239
	<i>Naucleopsis glabra</i> SPRUCE EX PITTIER	Tree	45	12	6270
	<i>Perebea mollis</i> (POEPP. & ENDL.) HUBER	Tree	45	14	6268
	<i>Pseudolmedia laevis</i> (RUIZ & PAV.) J. F. MACBR.	Tree	55	24	6323
	<i>Pseudolmedia macrophylla</i> TRÉCUL	Tree	10	2	6299
	<i>Sorocea muriculata</i> MIQ.	Tree	5	1	6192
	<i>Sorocea</i> sp.	Tree	60	22	6238
Myristicaceae	<i>Compsonera ulei</i> WARB.	Tree	15	3	6277
	<i>Iryanthera juruensis</i> WARB.	Tree	10	2	6255
	<i>Virola multinervia</i> DUCKE	Tree	25	6	6227
	<i>Virola</i> sp. 2	Tree	5	1	6257
	<i>Virola</i> sp.	Tree	5	1	6191
Myrtaceae	<i>Eugenia</i> sp. 1	Tree	5	1	
	<i>Eugenia acensis</i> McVAUGH	Tree	5	1	6300
	<i>Eugenia</i> sp. 2	Tree	65	26	6282
Nyctaginaceae	<i>Neea floribunda</i> POEPP. & ENDL.	Tree	55	23	
	<i>Neea</i> sp. 1	Tree	5	1	
	<i>Neea</i> sp. 2	Tree	15	3	6217
Olacaceae	<i>Aptandra tubicina</i> (POEPP.) BENTH. EX MIERS	Tree	10	2	6226
	<i>Chaunochiton</i> sp.	Tree	10	2	6195
	<i>Heisteria duckei</i> SLEUMER	Tree	10	2	6263
	<i>Minquartia guianensis</i> AUBL.	Tree	10	2	6205
Piperaceae	<i>Piper arboreum</i> AUBL.	Shrub	15	5	
	<i>Piper hispidinervum</i> C. DC.	Shrub	30	7	6289
	<i>Piper</i> sp. 1	Shrub	90	200	6273
	<i>Piper</i> sp. 2	Shrub	30	13	6315
	<i>Piper</i> sp. 3	Shrub	15	4	6314
Polygonaceae	<i>Triplaris</i> sp.	Tree	20	7	
Primulaceae	<i>Cybianthus guyanensis</i> subsp. <i>pseudoicacoreus</i> (MIQ.) PIPOLY	Tree	5	1	6251
Putranjivaceae	<i>Drypetes amazonica</i> STEYERM.	Tree	25	5	6308
Rubiaceae	<i>Alibertia claviflora</i> K. SCHUM.	Tree	5	1	6294
	<i>Alseis</i> sp.		5	1	6309

**Table 1.** Continuation.

Family	Species	Habit	Gaps occupied (%)	Abundance	Voucher A.S. Maranhó*
	<i>Amaioua guianensis</i> AUBL.	Tree	5	1	6311
	<i>Capirona decorticans</i> SPRUCE	Tree	5	4	6262
	<i>Faramea capillipes</i> MÜLL. ARG.	Tree	80	70	6229
	Rubiaceae		10	3	
	<i>Palicourea</i> sp.	Tree	20	8	6285
	<i>Psychotria hoffmannseggiana</i> (WILLD. EX SCHULT.) MÜLL. ARG.	Shrub	25	9	6197
	<i>Psychotria lupulina</i> BENTH.	Shrub	5	1	6190
	<i>Psychotria</i> sp. 1	Shrub	30	10	
	<i>Psychotria</i> sp. 2	Shrub	5	2	6201
Rutaceae	<i>Metrodorea flavida</i> K.KRAUSE	Tree	10	2	6261
Salicaceae	<i>Banara nitida</i> SPRUCE EX BENTH.	Tree	10	2	6194
	<i>Casearia gossypiosperma</i> BRIQ.	Tree	20	4	6202
Sapindaceae	<i>Allophylus pilosus</i> (J.F.MACBR.) A.H.GENTRY	Tree	30	22	6237
	<i>Talisia cerasina</i> (BENTH.) RADLK.	Shrub	10	3	6290
Sapotaceae	<i>Pouteria caimito</i> (RUIZ & PAV.) RADLK.	Tree	25	11	6272
	<i>Pouteria campanulata</i> BAEHNI	Tree	5	1	6204
	<i>Pouteria hispida</i> EYMA	Tree	15	4	6291
	<i>Pouteria</i> sp. 1	Tree	10	2	6298
	<i>Pouteria</i> sp. 2	Tree	5	1	
	<i>Pradosia atrovioleacea</i> DUCKE	Tree	25	8	6245
Siparunaceae	<i>Siparuna guianensis</i> AUBL.	Tree	65	31	6274
Solanaceae	<i>Brunfelsia grandiflora</i> D.DON	Shrub	25	5	6241
Thymelaeaceae	<i>Schoenobiblus peruviana</i> STANDL.	Tree	10	2	6301
Ulmaceae	<i>Ampelocera</i> sp.	Tree	15	3	
Urticaceae	<i>Pourouma cecropiifolia</i> MART.	Tree	5	1	6287
	<i>Pourouma guianensis</i> AUBL.	Tree	15	7	6292
	<i>Pourouma minor</i> BENOIST	Tree	50	11	6275
	<i>Urera</i> sp.		10	2	
Violaceae	<i>Amphirrhox</i> sp.	Tree	75	26	6271
	<i>Leonia glycyarpa</i> RUIZ & PAV.	Tree	30	10	6218
	<i>Rinoreocarpus ulei</i> (MELCH.) DUCKE	Tree	40	15	6222
Vochysiaceae	<i>Qualea grandiflora</i> MART.	Tree	15	6	6256
Total				1,218	

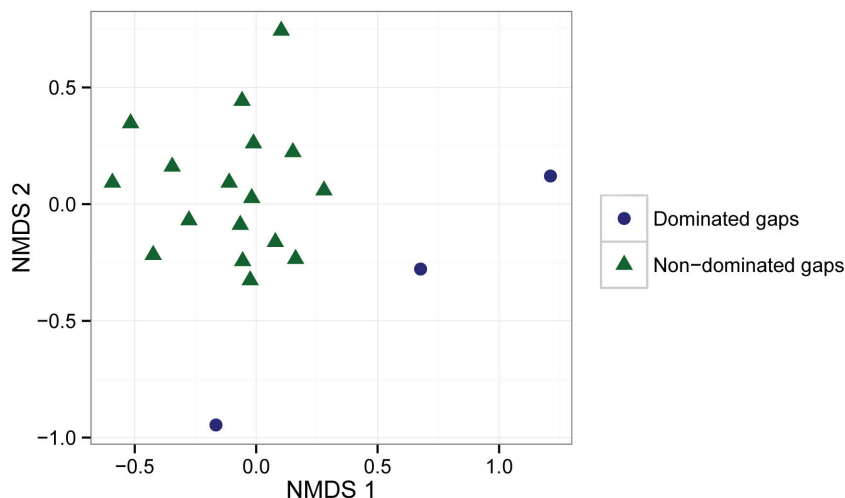
Note: (\*) initial code tipping specimens.

linear relationship (GLM,  $p < 0.001$ ;  $r^2=0.58$ ; Figure 4A). The local richness also decreased in areas dominated by Maranthaceae (GLM,  $p = 0.001$ ,  $r^2=0.48$ ; Figure 4B). Therefore, there

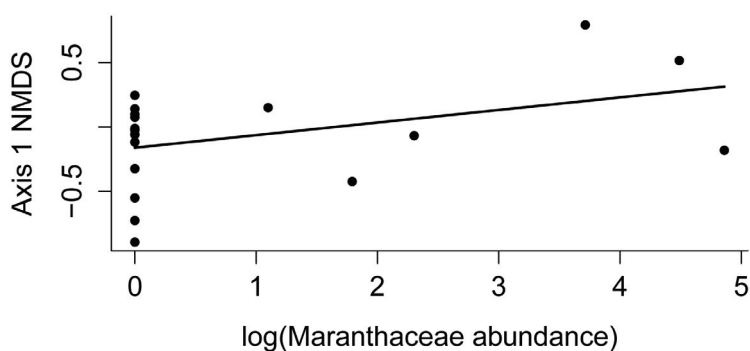
was an increase in dissimilarity between areas dominated and areas not dominated by Maranthaceae, as the abundance of individuals of this family increased. The decrease in woody

species abundance and richness in areas dominated by Maranthaceae can result from competition for space, explained by the extensive soil cover by Maranthaceae individuals (Figure 5).

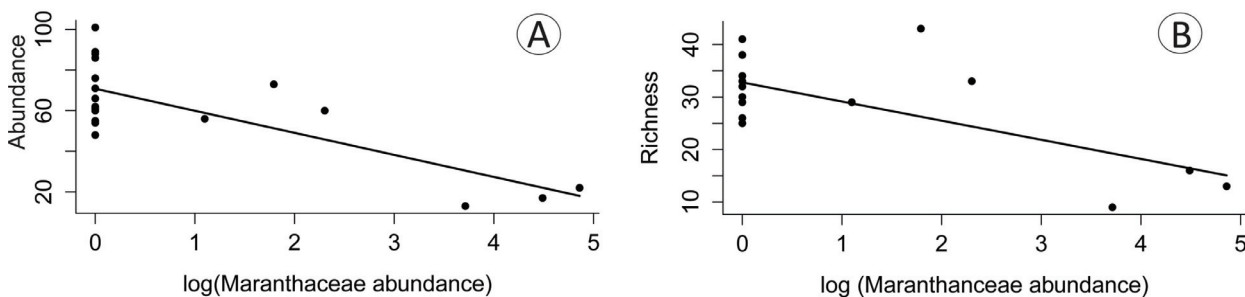
## Discussion



**Figure 2.** NMDS based on species composition of natural gaps, in Catuaba Experimental Farm, southwestern Amazonia, Brazil, grouped by dominance of Maranthaceae.



**Figure 3.** GLM between the axis 1 of the NMDS and Maranthaceae abundance ( $p = 0.07$ ) in gaps with or without dominance by Maranthaceae in Catuaba Experimental Farm, southwestern Amazonia, Brazil.



**Figure 4.** Species abundance (A) and richness (B) of woody plants as a function of Maranthaceae abundance in Maranthaceae dominated and non-dominated gaps in Catuaba Experimental Farm, southwestern Amazonia, Brazil.

In the present study, the overabundance of Maranthaceae decreased the abundance and richness of woody plants. Therefore, the gaps dominated by this family differed in species composition from the gaps not dominated by it. Similar results were presented by Lima and Moura (2006), who briefly reported the occurrence of two genera of Maranthaceae (*Calathea* and *Ctenanthe*) in forest gaps in the Atlantic Forest. These authors suggest that the wide leaves and the habit of forming dense clumps could produce a thick leaf litter, which could hinder the germination of other plants. Although not measured in the present study, we observed that where Maranthaceae are dense, their leaf litter also form a thick layer on soil, which could have the same effect as presented experimentally by Santos and Válio (2002) who showed the effect of litter on seedling recruitment and revealed the importance of film thickness in obstructing seedling emergence. According to Costa (2004), among the terrestrial herbs found in the *terra firme* forest in Central Amazonia, the family Maranthaceae was one of the richest in species and contributed strongly to soil cover when compared with the other herbs studied. In light of this, in gaps where Maranthaceae occurred and were dominant, such dense patches could form by the inhi-





**Figure 5.** Natural gaps with dominance by Maranthaceae, in the Catuaba Experimental Farm, southwestern Amazonia, Brazil. (A) A gap opening by a single tree fall; (B) a typical Maranthaceae; (C) a Maranthaceae dominated gap.

bition model of succession elucidated by Connell and Slatyer (1977). On the other hand, one must keep in mind that such colonization and dominance may also have occurred by mere stochastic events, such as the existence of seed dispersal in the time of gap opening and not because these species are more adapted than others to such gap environments.

We observed that Maranthaceae occurs in patches and seems to be favored by gap opening and the presence of watercourses. According to Costa and Magnusson (2002), gap opening and increase in light incidence can lead to an increase in the abundance of some herbaceous species, an effect observed in some Maranthaceae species. Costa *et al.* (2005) observed that the structure of the herbaceous community, in particular of Maranthaceae,

is also strongly affected by the type of soil and the presence of clay.

Such results could be caused by many factors, but the most probable cause of decrease in woody species abundance and richness is the competition for space. Gaps dominated by Maranthaceae had less physical space for other individuals' establishment. With such argument in mind, it would be desirable to understand and determine which the main factors controlling distribution and occurrence of Maranthaceae in these forests are.

### Conclusions

Maranthaceae overabundance decreased woody species abundance and richness significantly, which was corroborated by the dissimilarity among dominated and non-dominated gaps.

### References

- ACRE. 2006. *Zoneamento Ecológico-Econômico (ZEE) Fase II: documento síntese - Escala 1:250.000*. Rio Branco, SEMA, 356 p.
- BRANDANI, A.; HARTSHORN, G.S.; ORIAN, G.H. 1988. Internal heterogeneity of gaps and species richness in Costa Rican tropical wet forest. *Journal of Tropical Ecology*, **4**(2):99-119. <http://dx.doi.org/10.1017/S0266467400002625>
- CONNELL, J.H.; SLATYER, R.O. 1977. Mechanisms of succession in natural communities and their role in community stability and organization. *The American Naturalist*, **111**(982):1119-1144. <http://dx.doi.org/10.1086/283241>
- CONNELL, J.H. 1978. Diversity in tropical rain forests and coral reefs. *Science*, **199**(4335):1302-1310. <http://dx.doi.org/10.1126/science.199.4335.1302>
- COSTA, F.R.C. 2004. Structure and composition of the ground-herb community in a terra-firme Central Amazonian forest. *Acta Amazonica*, **34**(1):53-59. <http://dx.doi.org/10.1590/S0044-59672004000100007>

- COSTA, F.; MAGNUSSON, W.E. 2002. Selective logging effects on abundance, diversity, and composition of tropical understory herbs. *Ecological Applications*, **12**(3):807-819. [http://dx.doi.org/10.1890/1051-0761\(2002\)012\[0807:SLEOAD\]2.0.CO;2](http://dx.doi.org/10.1890/1051-0761(2002)012[0807:SLEOAD]2.0.CO;2)
- COSTA, F.R.C.; MAGNUSSON, W.E.; LUIZAO, R.C. 2005. Meso-scale distribution patterns of Amazonian understory herbs in relation to topography, soil and watersheds. *Journal of Ecology*, **93**(5):863-878. <http://dx.doi.org/10.1111/j.1365-2745.2005.01020.x>
- DENSLOW, J.S. 1980. Gap partitioning among tropical rainforest trees. *Biotropica*, **12**(2):47-55. <http://dx.doi.org/10.2307/2388156>
- DENSLOW, J.S. 1987. Tropical rainforest gaps and tree species diversity. *Annual Review of Ecology and Systematics*, **18**(1): 431-451. <http://dx.doi.org/10.1146/annurev.es.18.110187.002243>
- DUARTE, A.F. 2006. Aspectos da climatologia do Acre, Brasil, com base no intervalo 1971-2000. *Revista Brasileira de Meteorologia*, **21**(3):308-317.
- GARROT, R.A.; WHITE, P.J.; WHITE, C.A.V. 1993. Overabundance: an issue conservation biologists? *Conservation Biology*, **7**(4):946-949. <http://dx.doi.org/10.1046/j.1523-1739.1993.740946.x>
- GENTRY, A.H.; EMMONS, L.H. 1987. Geographical variation in fertility, phenology, and composition of the understory of neotropical forests. *Biotropica*, **19**(3):216-227. <http://dx.doi.org/10.2307/2388339>
- HUBBELL, S.P.; FOSTER, R.B.; O'BRIEN, S.T.; HARMS, K.E.; CONDIT, R.; WESCHLER, B.; WRIGHT, S.J.; LOO DE LAO, S. 1999. Light-gap disturbances, recruitment limitation, and tree diversity in a Neotropical Forest. *Science*, **283**(5401):554-557. <http://dx.doi.org/10.1126/science.283.5401.554>
- LIMA, R.A.F.; MOURA, L.C. 2008. Gap disturbance regime and composition in the Atlantic Montane Rain Forest: the influence of topography. *Plant Ecology*, **197**(2):239-253. <http://dx.doi.org/10.1007/s11258-007-9374-x>
- LIMA, R.A.F.; ROTHER, D.C.; MULER, A.E.; LEPSCH, I.F.; RODRIGUES, R.R. 2012. Bamboo overabundance alters forest structure and dynamics in the Atlantic forest hotspot. *Biological Conservation*, **147**(1):32-39. <http://dx.doi.org/10.1016/j.biocon.2012.01.015>
- MAGNUSSON, W.E.; LIMA, A.P.; LUIZÃO, R.; LUIZÃO, F.; COSTA, F.R.C.; CASTILHO, C.V.; KINUPP, V.F. 2005. RAPELD: a modification of the Gentry method for biodiversity surveys in long-term ecological research sites. *Biota Neotropica*, **5**(2):1-6. <http://dx.doi.org/10.1590/S1676-06032005000300002>
- OKSANEN, J.F.; BLANCHET, G.; KINDT, R.; LEGENDRE, P.; MINCHIN, P.R.; O'HARA, R.B.; SIMPSON, G.L.; SOLYMOS, P.; STEVENS, M.H.H.; WAGNER, H. 2013. Vegan: Community Ecology Package. R package version 2.0-8. Available at: <http://CRAN.R-project.org/package=vegan>. Accessed on: 11/10/2014.
- POULSEN, A.D.; BALSLEV, H. 1991. Abundance and cover of ground herbs in an Amazonian rain forest. *Journal of Vegetation Science*, **2**(3):315-322. <http://dx.doi.org/10.2307/3235922>
- R CORE TEAM. 2013. R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. Available at: <http://www.R-project.org/>. Accessed on: 11/10/2014.
- RIBEIRO, S.C.; BOTELHO, S.A.; FONTES, M.A.L.; GARCIA, P.O.; ALMEIDA, H. de S. 2013. Regeneração natural em áreas desmatadas e dominadas por *Pteridium aquilinum* (L.) Kuhn. na Serra da Mantiqueira. *Cerne*, **19**(1):65-76. <http://dx.doi.org/10.1590/S0104-77602013000100009>
- RUNKLE, J.R. 1981. Gap regeneration in some old-growth forests of the Eastern United States. *Ecology*, **62**(4):1041-1051. <http://dx.doi.org/10.2307/1937003>
- RUNKLE, J.R. 1992. *Guidelines and sample protocol for sampling forest gaps*. Portland, Department of Agriculture, Forest Service, Pacific Northwest Research Station, 44 p.
- SANTOS, S.L. dos; VÁLIO, I.F.M. 2002. Litter accumulation and its effect on seedling recruitment in a Southeast Brazilian Tropical Forest. *Revista Brasileira de Botânica*, **25**(1):89-92. <http://dx.doi.org/10.1590/s0100-84042002000100011>
- SILVEIRA, M. 2005. *A floresta aberta com bambu no sudoeste da Amazônia: padrões e processos em múltiplas escalas*. Rio Branco, EDUFAC, 157 p.
- SCHLIEMANN, S.A.; BOCKHEIM, J.G. 2011. Methods for studying treefall gaps: a review. *Forest Ecology and Management*, **261**(7):1143-1151. <http://dx.doi.org/10.1016/j.foreco.2011.01.011>
- SCHNEIDER, L.C. 2006. Invasive species and land-use: the effect of land management practices on bracken fern invasion in the region of Calakmul, Mexico. *Journal of Latin American Geography*, **5**(2):92-107. <http://dx.doi.org/10.1353/lag.2006.0028>
- SOUSA, E. dos S.; SALIMON, C. I.; VICTORIA, R.L.; KRUSHE, A.; ALIN, S.; LEITE, N.K. 2008. Dissolved inorganic carbon and pCO<sub>2</sub> in two small streams draining different soil types in Southwestern Amazonia, Brazil. *Revista Ambiente & Água*, **3**(2):37-50. <http://dx.doi.org/10.4136/ambi-agua.51>
- VEBLEN, T.T. 1989. Tree regeneration responses to gaps along a Transandean gradient. *Ecology*, **70**(3):541-543. <http://dx.doi.org/10.2307/1940197>
- WRIGHT, S.J. 2002. Plant diversity in tropical forests: a review of mechanisms of species coexistence. *Oecologia*, **130**(1):1-14. <http://dx.doi.org/10.1007/s004420100809>

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